# Determination of Ultratrace Polychlorinated Dibenzo-p-Dioxins and Dibenzofurans Using GC/MS/MS 


#### Abstract

Polychlorinated dibenzo-p-dioxins (PCDDs) and polychlorinated dibenzofurans (PCDFs) are highly toxic persistent organic pollutants (POPs). Analysis of these toxic PCDD/F congeners is very challenging because they are difficult to detect at ultratrace levels in complex samples. This study developed a gas chromatography triple quadrupole mass spectrometry (GC/MS/MS) method for the analysis of seventeen 2,3,7,8-substituted PCDD/F congeners. It was performed using an Agilent 7010 GC/MS/MS with a novel high-efficiency electron ionization source that can detect trace and ultratrace levels of analytes with higher sensitivity and confidence.

Incineration is the main source of dioxins in the environment. Therefore, the analysis of dioxins in waste incineration fly ash samples is of great significance for controlling the emission of dioxins. Six fly ash samples with varied concentrations between 2.1 and 32.6 pg TEQ/g were analyzed using GC/MS/MS. The GC/MS/MS results were in agreement with values obtained using GC/HRMS. The method was also validated through the analysis of a certified reference material of fish tissue with five injections. For all of the congeners, the average results from GC/MS/MS were in the range of the certified reference values. The relative standard deviations (RSDs) of all the congeners were less than 10.0\%. Therefore, this GC/MS/MS method provides a viable and economical alternative to the GC/HRMS method.


## Introduction

PCDD/Fs are of great concern because of their toxicity and persistence in the environment. ${ }^{1}$ Analysis of these toxic PCDD/F congeners is challenging because they are difficult to detect at ultratrace levels in complex environmental samples. GC/MS/MS has a specific multiple reaction monitoring (MRM) mode that creates specific fragmentation of the PCDD/Fs. This selective reaction can decrease the interference in mass chromatograms and improve sensitivity. Because of this, gas chromatography coupled with tandem mass spectrometry has been applied to the analysis of PCDD/Fs. The European Union (EU) has instituted regulation (709/2014) governing the levels of PCDDs and PCDFs in food and feed that enables the use of GC/MS/MS systems in confirmatory testing for compliance with EU MLs. This means that triple quadrupole mass spectrometers could provide performance similar to HRMS systems. ${ }^{2}$ This Application Note describes the sensitivity, selectivity, and precision of the methods for analyzing 17 toxic PCDD/Fs using an Agilent 7010 triple quadrupole GC/MS system. Table1 shows the specific compounds, along with the assigned Toxicity Equivalency Factor (TEF), International Toxicity Equivalency Factor (I-TEF), and World Health Organization TEF ( $\mathrm{WHO}_{2005}-$ TEF) to calculate toxic equivalency (TEQ). This Application Note also shows that the GC/MS/MS results agree with values obtained using GC/HRMS, thus providing a viable and economical alternative to the GC/HRMS approach.

## Experimental

## Reagents and standards

Residue grade $n$-hexane, dichloromethane, and toluene were purchased from J.T. Baker (Phillipsburg, $\mathrm{NJ}, \mathrm{USA}$ ). Standard solutions for 2,3,7,8-PCDD/Fs specified by EPA Method 1613, including those for EPA-1613 CVS, LCS, ISS, and certified reference material WMF-01 (reference fish tissue) were supplied by Wellington Laboratories Inc (Ontario, Canada).

## Instruments

The analyses were performed on an Agilent 7890 GC and an Agilent 7010 triple quadrupole GC/MS system. Table 2 lists the instrument conditions. The methods used MRM mode for data acquisition. For each target, two

Table 1. Various toxic equivalent factors (TEF) of PCDD/Fs. ${ }^{3}$

| PCDD/Fs | I-TEF | WHO $_{2005}$-TEF |
| :--- | :--- | :--- |
| 2,3,7,8-TCDD | 1 | 1 |
| $1,2,3,7,8-P e C D D$ | 0.5 | 1 |
| $1,2,3,4,7,8-$ HxCDD | 0.1 | 0.1 |
| $1,2,3,6,7,8-H x C D D$ | 0.1 | 0.1 |
| $1,2,3,7,8,9-H x C D D$ | 0.1 | 0.1 |
| $1,2,3,4,6,7,8-H p C D D$ | 0.01 | 0.01 |
| OCDD | 0.001 | 0.0003 |
| $2,3,7,8-$ TCDF | 0.1 | 0.1 |
| $1,2,3,7,8-P e C D F$ | 0.05 | 0.03 |
| $2,3,4,7,8-P e C D F$ | 0.5 | 0.3 |
| $1,2,3,4,7,8-H x C D F$ | 0.1 | 0.1 |
| $1,2,3,6,7,8-H x C D F$ | 0.1 | 0.1 |
| $1,2,3,7,8,9-H x C D F$ | 0.1 | 0.1 |
| $2,3,4,6,7,8-H x C D F$ | 0.1 | 0.1 |
| $1,2,3,4,6,7,8-H p C D F$ | 0.01 | 0.01 |
| $1,2,3,4,7,8,9-H p C D F$ | 0.01 | 0.01 |
| $0 C D F$ | 0.001 | 0.0003 | specific precursor ions as well as two corresponding product ions and collision energies were adapted from Agilent Food and Feed Analyzer. ${ }^{4}$ Table 3 gives a full list of MRM transitions.

Table 2. Instrument conditions.

| GC Conditions |  |
| :---: | :---: |
| Column | Agilent J\&W DB-5ms UI, $60 \mathrm{~m} \times 0.25 \mathrm{~mm}, 0.25 \mu \mathrm{~m}$ |
| Injection Volume | $1 \mu \mathrm{~L}$ |
| Oven Temperature | $150^{\circ} \mathrm{C}$ held for 3 minutes, at $20^{\circ} \mathrm{C} / \mathrm{min}$ to $230^{\circ} \mathrm{C}$ held for 18 minutes, at $5^{\circ} \mathrm{C} / \mathrm{min}$ to $235^{\circ} \mathrm{C}$ held for 10 minutes, at $4^{\circ} \mathrm{C} / \mathrm{min}$ to $320^{\circ} \mathrm{C}$ held for 1 minute |
| Injection Mode | Splitless, purge on after 1.5 minutes |
| Injection Port Temperature | $290{ }^{\circ} \mathrm{C}$ |
| Carrier Gas | Helium |
| Flow Rate | $1.0 \mathrm{~mL} / \mathrm{min}$ |
| MS Conditions |  |
| Operation Mode | Electron ionization (EI), MRM |
| Ionization Voltage | 70 eV |
| Ion Source Temperature | $320{ }^{\circ} \mathrm{C}$ |
| Interface Temperature | $320{ }^{\circ} \mathrm{C}$ |
| Quadrupole Temperature | $150{ }^{\circ} \mathrm{C}$ |
| Solvent Delay | 10 minutes |
| MS1 Resolution | Unit |
| MS2 Resolution | Unit |
| Collision Cell Gas Flows | Nitrogen at $1.5 \mathrm{~mL} / \mathrm{min}$, helium at $4.0 \mathrm{~mL} / \mathrm{min}$ |

Table 3. Main parameters for MS/MS analysis of PCDD/Fs.

| Compound | Precursor ion (m/z) | Product ion (m/z) | Collision energy (eV) |
| :---: | :---: | :---: | :---: |
| TCDF | 303.9 | 240.9 | 40 |
|  | 305.9 | 242.9 | 40 |
| ${ }^{13} \mathrm{C}_{12}$-TCDF | 315.9 | 251.9 | 40 |
|  | 317.9 | 253.9 | 40 |
| TCDD | 319.9 | 256.9 | 26 |
|  | 321.9 | 258.9 | 26 |
| ${ }^{13} \mathrm{C}_{12}$-TCDD | 331.9 | 267.9 | 26 |
|  | 333.9 | 269.9 | 26 |
| PeCDF | 337.9 | 274.9 | 40 |
|  | 339.9 | 276.9 | 40 |
| ${ }^{13} \mathrm{C}_{12}$-PeCDF | 349.9 | 285.9 | 40 |
|  | 351.9 | 287.9 | 40 |
| PeCDD | 353.9 | 290.9 | 26 |
|  | 355.9 | 292.9 | 26 |
| ${ }^{13} \mathrm{C}_{12}$-PeCDD | 365.9 | 301.9 | 26 |
|  | 367.9 | 303.9 | 26 |
| HxCDF | 373.8 | 310.9 | 40 |
|  | 375.8 | 312.9 | 40 |
| ${ }^{13} \mathrm{C}_{12}-\mathrm{HxCDF}$ | 385.8 | 321.9 | 40 |
|  | 387.8 | 323.9 | 40 |


| Compound | Precursor ion (m/z) | Product ion (m/z) | Collision energy (eV) |
| :---: | :---: | :---: | :---: |
| HxCDD | 389.8 | 326.9 | 26 |
|  | 391.8 | 328.8 | 25 |
| ${ }^{13} \mathrm{C}_{12}-\mathrm{HxCDD}$ | 401.8 | 337.9 | 26 |
|  | 403.8 | 339.9 | 25 |
| HpCDF | 407.8 | 344.8 | 40 |
|  | 409.8 | 346.8 | 40 |
| ${ }^{13} \mathrm{C}_{12}$ - HpCDF | 419.8 | 355.8 | 40 |
|  | 421.8 | 357.8 | 40 |
| HpCDD | 423.8 | 360.8 | 24 |
|  | 425.8 | 362.9 | 24 |
| ${ }^{13} \mathrm{C}_{12}$ - HpCDD | 435.8 | 371.8 | 24 |
|  | 437.8 | 373.8 | 24 |
| OCDF | 441.7 | 378.8 | 40 |
|  | 443.7 | 380.8 | 40 |
| ${ }^{13} \mathrm{C}_{12}$-OCDF | 455.8 | 391.8 | 40 |
|  | 453.8 | 389.8 | 40 |
| OCDD | 457.7 | 394.8 | 24 |
|  | 459.7 | 396.8 | 24 |
| ${ }^{13} \mathrm{C}_{12}$-OCDD | 469.7 | 405.8 | 24 |
|  | 471.7 | 407.8 | 24 |

## Sample preparation

Samples required elaborate extraction and cleanup procedures before instrumental analysis. Before extraction, samples were spiked with known amounts of the ${ }^{13} \mathrm{C}_{12}$-labeled standards for EPA-1613 LCS, then equilibrated for 12 hours. After that, samples were extracted with $n$-hexane and dichloromethane (1:1, V/V), using Soxhlet or ASE. After concentration, the extractions were cleaned by three columns: an acidified silica gel column, a multilayered acid/base/ $\mathrm{AgNO}_{3} /$ neutral silica gel, and an active carbon column. The final extracts were eluted with toluene, then reduced under a gentle stream of purified nitrogen to an appropriate volume. For recovery quantification, the ${ }^{13} \mathrm{C}_{12}$-labeled standards for EPA-1613 ISS were added immediately before instrumental analysis. Figure 1 shows a flow diagram summarizing the sample preparation steps.


Figure 1. Flow diagram of the sample extraction and cleanup procedures.

## Results and discussion

## The separation of the dioxin/furan isomers

Figure 2 shows the MRM chromatograms for 17 PCDD/F congeners, with an analysis time of 58 minutes. It shows the excellent separation of 17 PCDD/F congeners and it also zooms in on the hexa-dioxin/furan isomers that were difficult to separate.

## Calibration and average relative response factor

The optimized GC/MS/MS method was applied to analyze the calibration standard solutions EPA-1613 CVS (CS1 to CS5) (Table 4). According to EPA Method 1613,5 the relative response factor (RRF) of each individual 2,3,7,8-chloro-substituted PCDD/F congener was obtained from a five-point calibration curve. The RSDs of all the congeners were below $3.0 \%$, which fully complies with the EPA's requirement of less than 15\% (Table 5).


Figure 2. MRM chromatograms of 17 PCDD/F congeners.

Table 4. Concentrations of congeners in calibration solutions ( $\mathrm{ng} / \mathrm{mL}$ ).

| Native PCDD/Fs | 1613 CS1 | 1613 CS2 | 1613 CS3 | 1613 CS4 | 1613 CS5 |
| :--- | :---: | :---: | :---: | :---: | :---: |
| 2,3,7,8-TCDF | 0.5 | 2 | 10 | 40 | 200 |
| $1,2,3,7,8-P e C D F$ | 2.5 | 10 | 50 | 200 | 1,000 |
| $2,3,4,7,8-P e C D F$ | 2.5 | 10 | 50 | 200 | 1,000 |
| $1,2,3,4,7,8-H x C D F$ | 2.5 | 10 | 50 | 200 | 1,000 |
| $1,2,3,6,7,8-H x C D F$ | 2.5 | 10 | 50 | 200 | 1,000 |
| $2,3,4,6,7,8-H x C D F$ | 2.5 | 10 | 50 | 200 | 1,000 |
| $1,2,3,7,8,9-H x C D F$ | 2.5 | 10 | 50 | 200 | 1,000 |
| $1,2,3,4,6,7,8-H p C D F$ | 2.5 | 10 | 50 | 200 | 1,000 |
| $1,2,3,4,7,8,9-H p C D F$ | 2.5 | 10 | 50 | 200 | 1,000 |
| 0CDF | 5.0 | 20 | 100 | 400 | 2,000 |
| $2,3,7,8-T C D D$ | 0.5 | 2 | 10 | 40 | 200 |
| $1,2,3,7,8-P e C D D$ | 2.5 | 10 | 50 | 200 | 1,000 |
| $1,2,3,4,7,8-H x C D D$ | 2.5 | 10 | 50 | 200 | 1,000 |
| $1,2,3,6,7,8-H x C D D$ | 2.5 | 10 | 50 | 200 | 1,000 |
| $1,2,3,7,8,9-H x C D D$ | 2.5 | 10 | 50 | 200 | 1,000 |
| $1,2,3,4,6,7,8-H p C D D$ | 2.5 | 10 | 50 | 200 | 1,000 |
| $0 C D D$ | 5.0 | 20 | 100 | 400 | 2,000 |

## Limits of detection (LODs)

The LODs for PCDD/Fs were determined from seven replicate analyses of a standard solution containing low concentrations (diluted 10 times of CS1) of the PCDD/Fs according to the US EPA method. The obtained LODs ranged from 0.008 to $0.08 \mathrm{pg} / \mu \mathrm{L}$ for the 17 PCDD/F congeners, which suggested that the GC/MS/MS method was sensitive enough for ultratrace analysis of PCDD/Fs. Table 5 shows the results. Figure 3 shows MRM chromatograms of 17 PCDD and PCDF congeners. (Diluted 10 times of CS1, from 0.05 to $0.5 \mathrm{pg} / \mu \mathrm{L})$.

Table 5. Average RRFs for individual congener calibrations and their corresponding LODs

| Compound name | Average RRF | RSD (\%) | LOD (pg/ $\mu \mathrm{L}$ ) |
| :--- | :---: | :---: | :---: |
| 2,3,7,8-TCDF | 1.07 | 1.07 | 0.01 |
| 1,2,3,7,8-PeCDF | 1.03 | 1.27 | 0.049 |
| 2,3,4,7,8-PeCDF | 1.06 | 1.39 | 0.039 |
| $1,2,3,4,7,8-H x C D F$ | 1.02 | 1.81 | 0.04 |
| $1,2,3,6,7,8-H x C D F$ | 1.01 | 1.89 | 0.04 |
| $2,3,4,6,7,8-H x C D F$ | 1.08 | 1.36 | 0.04 |
| $1,2,3,7,8,9-H x C D F$ | 0.95 | 1.46 | 0.04 |
| $1,2,3,4,6,7,8-H p C D F$ | 0.99 | 1.56 | 0.04 |
| $1,2,3,4,7,8,9-H p C D F$ | 0.98 | 1.17 | 0.04 |
| OCDF | 1.56 | 1.48 | 0.08 |
| $2,3,7,8-T C D D$ | 1.15 | 0.86 | 0.008 |
| $1,2,3,7,8-P e C D D$ | 1.07 | 1.96 | 0.04 |
| $1,2,3,4,7,8-H x C D D$ | 1.05 | 2.04 | 0.04 |
| $1,2,3,6,7,8-H x C D D$ | 0.96 | 1.39 | 0.05 |
| $1,2,3,7,8,9-H x C D D$ | 0.88 | 2.46 | 0.04 |
| $1,2,3,4,6,7,8-H p C D D$ | 0.98 | 1.18 | 0.04 |
| OCDD | 1.00 | 0.95 | 0.08 |





45.546 .046 .547 .0


$35.0 \quad 35.5 \quad 36.0$


$38.539 .0 \quad 39.5$


$46.0 \quad 46.5 \quad 47.0 \quad 47.5$

46.547 .047 .548 .0

47.047 .548 .048 .5

Figure 3. MRM chromatograms of 17 PCDD and PCDF congeners (diluted 10 times of CS1, from 0.05 to $0.5 \mathrm{pg} / \mathrm{\mu L}$ ).

## The evaluation of the proposed analytical method

To evaluate the performance of the proposed GC/MS/MS method, it was applied to the analysis of certified reference material (CRM) fish tissue with five replications. Table 6 shows that the concentrations of PCDD/F congeners in fish tissue ranged from 0.23 to $13.6 \mathrm{pg} / \mathrm{g}$, and the average results of all the congeners obtained from GC/MS/MS were in the range of the certified reference values. The total I-TEQ result of GC/MS/MS was $19.92 \mathrm{pg} / \mathrm{g}$, which was close to the certified reference value $19.81 \mathrm{pg} / \mathrm{g}$. The RSD of the five injections was less than 10\%. The average ion abundance ratio for 17 PCDD/F congeners of CRM (Figure 4), which are all within a $\pm 15 \%$ window around the average ion abundance of CS1 to CS5, meet the requirement of EPA 1613. Overall, the proposed analytical method showed good accuracy and precision.

Table 6. Analysis of PCDD/Fs in certified reference material (WMF-01).

|  | Certified Reference Value (pg/g) | Analyzed Value ( $\mathrm{n}=5$ ) |  |
| :---: | :---: | :---: | :---: |
|  |  | Average (pg/g) | RSD (\%) |
| 2,3,7,8-TCDF | $13.1 \pm 4.9$ | 12.97 | 1.0 |
| 1,2,3,7,8-PeCDF | $1.53 \pm 1.4$ | 1.34 | 8.6 |
| 2,3,4,7,8-PeCDF | $7.15 \pm 2.2$ | 6.43 | 3.3 |
| 1,2,3,4,7,8-HxCDF | $0.86 \pm 1.0$ | 1.01 | 6.5 |
| 1,2,3,6,7,8-HxCDF | $0.51 \pm 0.7$ | 0.62 | 6.0 |
| 2,3,4,6,7,8-HxCDF | $0.68 \pm 1.2$ | 0.67 | 3.9 |
| 1,2,3,7,8,9-HxCDF | $0.25 \pm 0.4$ | 0.26 | 8.8 |
| 1,2,3,4,6,7,8-HpCDF | $1.01 \pm 1.9$ | 2.76 | 2.8 |
| 1,2,3,4,7,8,9-HpCDF | $0.30 \pm 0.5$ | 0.61 | 3.8 |
| OCDF | $1.38 \pm 2.1$ | 2.94 | 8.7 |
| 2,3,7,8-TCDD | $13.1 \pm 4.4$ | 13.6 | 2.4 |
| 1,2,3,7,8-PeCDD | $2.72 \pm 1.3$ | 2.61 | 2.9 |
| 1,2,3,4,7,8-HxCDD | $0.22 \pm 0.3$ | 0.27 | 7.4 |
| 1,2,3,6,7,8-HxCDD | $0.88 \pm 0.4$ | 0.81 | 9.2 |
| 1,2,3,7,8,9-HxCDD | $0.27 \pm 0.4$ | 0.23 | 7.6 |
| 1,2,3,4,6,7,8-HpCDD | $0.59 \pm 0.7$ | 0.65 | 6.5 |
| OCDD | $3.91 \pm 6.2$ | 2.01 | 7.9 |
| Total I-TEQ | 19.81 | 19.92 |  |



Figure 4. Comparative average ion abundance ratio for 17 PCDD/F congeners of CS1-CS5 and CRM.

## The comparison of GC/HRMS and GC/MS/MS for the analysis of PCDD/Fs from fly ash samples

Six fly ash samples were extracted and analyzed using a GC/HRMS. The same sample vials were then transferred to the GC/MS/MS and reanalyzed. Figure 5 shows the comparative sample results (total I-TEQ) of the two sets of measurements between the results obtained by the GC/HRMS and GC/MS/MS analysis of six fly ash samples with varied concentrations between 2.1 and 32.6 pg TEQ/g. The GC/MS/MS results were in agreement with values obtained using GC/HRMS.

## Conclusion

The Agilent 7010 GC/MS/MS system provided reproducible and sensitive detection of 17 toxic PCDD/F congeners. The proposed method was applied to the analysis of certified reference materials to demonstrate its suitability. The results from GC/MS/MS were close to the certified reference values. Comparison of analytical results by GC/HRMS and GC/MS/MS indicated the suitability of the 7010 GC/MS/MS system.


Figure 5. Comparative sample results (total I-TEQ) of the two sets of measurements by the GC/HRMS and GC/MS/MS analyses.

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